






Cross-jurisdictional larval supply essential for eastern Australian spanner crabs (*Ranina ranina*)

Hayden T. Schilling^{A,B,C,*} , Daniel E. Hewitt^B , Neil Malan^B , Matthew D. Taylor^{B,C}  and Daniel D. Johnson^C 

For full list of author affiliations and declarations see end of paper

***Correspondence to:**

Hayden T. Schilling
Port Stephens Fisheries Institute, New South
Wales Department of Primary Industries,
Locked Bag 1, Nelson Bay, NSW 2315,
Australia
Email: hayden.schilling@dpi.nsw.gov.au

Handling Editor:

Jill Chiu

Received: 10 December 2021

Accepted: 15 August 2022

Published: 9 September 2022

Cite this:

Schilling HT *et al.* (2022)
Marine and Freshwater Research
doi: [10.1071/MF21348](https://doi.org/10.1071/MF21348)

© 2022 The Author(s) (or their employer(s)). Published by CSIRO Publishing.
This is an open access article distributed under the Creative Commons Attribution-NonCommercial 4.0 International License (CC BY-NC).

OPEN ACCESS

ABSTRACT

Context. The spanner crab (*Ranina ranina*) stock of eastern Australia is distributed across two state jurisdictions and, as a non-migratory species with a pelagic larval phase, connectivity within this stock is likely to occur by larval dispersal, driven by ocean currents. **Aims.** To understand connectivity and patterns of larval supply in the eastern Australian spanner crab stock. **Methods.** Lagrangian particle tracking methods were used to simulate larval transport around the key spanner crab fishing regions in eastern Australia. **Key results.** Spawning off central Queensland (Qld) supplies a large proportion of recruits, supporting both the Qld and New South Wales (NSW) fisheries. Lagged larval settlement showed significant correlations to catch-per-unit-effort and the proportion of total harvest taken within the NSW fishery, providing evidence to suggest that the NSW fishery may be reliant on spawning activity in Queensland. **Conclusions.** The Qld and NSW fisheries are highly connected and the broad-scale patterns identified by the current modelling approach could provide an indicator of potentially good or bad recruitment years, particularly as finer resolution, and refined reproductive biology knowledge on spanner crabs becomes available. **Implications.** The Qld and NSW fisheries are highly connected with a source–sink structure and it is recommended that a co-management strategy be adopted.

Keywords: connectivity, East Australian Current, Lagrangian, larval subsidy, modelling, particle tracking, pelagic larval duration, temperature dependent growth.

Introduction

Variability in the productivity of marine fisheries is primarily driven by dynamics of the exploited stock (Anderson *et al.* 2008). It is increasingly being recognised that environmental conditions have a significant influence on larval survival, dispersal and recruitment to fisheries (Dixon *et al.* 1999; Szuwalski *et al.* 2015). However, the mechanistic relationships between environmental and population dynamics are often poorly understood. This limits our ability to model variation in productivity, which has implications for the sustainable management of exploited stocks (Yatsu *et al.* 2005; Kurota *et al.* 2020).

For many exploited marine species, dispersal is driven by ocean currents during a pelagic larval stage. For such species, connectivity patterns are defined during the larval stage and can cross the jurisdictional boundaries within which exploited species are typically managed (Everett *et al.* 2017). It is difficult to convincingly measure larval connectivity in oceanic systems (White *et al.* 2019). Genetic techniques are often employed to examine genetic transfer between regions (Lowe and Allendorf 2010; Saenz-Agudelo *et al.* 2011). Unfortunately, genetic techniques provide little insight into the dispersal mechanisms of the larvae, which are often driven by oceanography, and genetic connectivity can be driven by the movement of few individuals (Cowen *et al.* 2007). Provided appropriate models are available, particle tracking simulations can show the influence of oceanographic processes, including dispersal mechanisms and pathways.

The coupling of hydrodynamic models with particle tracking simulations, which can incorporate species-specific biological rates (e.g. growth, mortality), allows the

re-creation of past dispersal events (Gallego *et al.* 2007; Hinrichsen *et al.* 2011). Examining these past dispersal events alongside oceanographic conditions can show both patterns and drivers of connectivity and recruitment. Particle tracking techniques have been applied successfully in many investigations at scales from small embayments (Norrie *et al.* 2020) to global oceans (Doblin and van Sebille 2016). Further, these techniques have been applied to a variety of taxa including microbes (Doblin and van Sebille 2016), kelp (Coleman *et al.* 2013), invertebrates (Everett *et al.* 2017; Cetina-Heredia *et al.* 2019) and fish (Schilling *et al.* 2020). By modelling larval dispersal of fished species, it is possible to investigate the influence of oceanographic conditions on interannual fluctuations in recruitment and the impacts on variability in fisheries productivity.

Spanner crabs (*Ranina ranina*) are large brachyurans found throughout the Indo-Pacific region. Like many marine decapods, spanner crabs have a two-phase life cycle with a pelagic larval phase and a benthic post-settlement, non-migratory juvenile–adult phase (Brown *et al.* 2008; Kapur *et al.* 2019). The species is commercially harvested throughout much of the Indo-Pacific (Kapur *et al.* 2019), with the largest fishery occurring in the waters off eastern Australia (Scandol and Kennelly 2002; O'Neill *et al.* 2010). The stock in Australia is managed independently by the states of New South Wales (NSW) and Queensland (Qld), with the latter accounting for most of the harvest (e.g. ~80% in 2017).

The combined commercial harvest of spanner crabs from Qld and NSW declined from >4000 to <1000 tonnes between 1994 and 2019 calendar years. The reduction in the overall catch is partially attributed to large reductions in effort owing to management changes in both jurisdictions, reducing the overall fleet size. The Qld fishery is recognised to have declined substantially since the mid-1990s, with reduced harvests and catch rates in both the fishery and fishery-independent surveys, and overall spanner crab biomass is currently estimated to be below target reference levels (Queensland Government 2021). Within the NSW fishery, which accounts for ~20% of the overall harvest in this stock (Roelofs *et al.* 2021), the catch has also declined since the 1990s (from ~465 to ~110 tonnes in 2018). However, the main performance indicators (standardised catch rates) for the NSW fishery were tracking above long-term averages from 2015 to 2019, although declines are evident in recent years (Roelofs *et al.* 2021; Supplementary Fig. S1). The total allowable catch in both NSW (169 to 140 tonnes) and Qld (1631 to 847 tonnes) has been reduced in recent years to restrict fishing mortality, increase protection of the spawning stock, and halt the decline in spanner crab stocks (NSW Total Allowable Fishing Committee 2020; Queensland Government 2021). Within NSW, in 2015, a peak in sublegal-size crabs was observed in the independent survey, which corresponded to a peak in commercial catch-per-unit-effort (CPUE) the following year (Johnson 2021).

The peak in both subadult and subsequent legal-size crabs was likely to be driven by favourable environmental conditions increasing the catchability of crabs (Spencer *et al.* 2017), or a strong recruitment event. However, source–sink dynamics and larval dispersal patterns, which may provide insight into recruitment, are unclear for the species.

In general, spanner crabs spawn throughout the fished areas of the species distribution (Kennelly and Watkins 1994; Brown *et al.* 2008; Zones 2–7 in Fig. 1). The spawning biomass is protected by temporal closures, in both NSW and Qld, which align with periods of peak reproductive activity (November–December–January), as well as minimum-size limits (NSW: 9.3-cm orbital carapace length, CL; Qld: 10-cm rostral CL; these are approximately equal-sized crabs but measured differently) that aim to allow mature individuals to spawn at least once before recruiting to the fishery, while helping keep a higher proportion of females in the population. In eastern Australia, regionally variable environmental conditions, including temperature, chlorophyll-*a*, bottom current velocity, and wind-driven upwelling, have large influences on catch rates (Spencer *et al.* 2017, 2019a, 2019b, 2019c). However, little is known about larval connectivity, and resulting recruitment. This study, therefore, aimed to use particle tracking simulations to (1) understand the patterns of larval dispersal and connectivity in the eastern Australian spanner crab stock; and, (2) investigate correlations between predicted settlement rates and NSW commercial CPUE for the species.

Materials and methods

Study region, fisheries, and biology

The eastern Australian spanner crab stock is primarily located between 23 and 29°S, with a very small amount of fishing occurring north and south of the main area (Fig. 1). Fishing occurs within six main zones, with Zones 2–6 being managed by Queensland Department of Agriculture and Fisheries (QDAF) and Zone 7 managed by New South Wales Department of Primary Industries – Fisheries (DPIF; Fig. 1). Crabs are harvested using flat dillies, designed to entangle rather than trap, deployed in depths between 30 and 80 m (Dichmont and Brown 2010). In NSW, following the introduction of daily reporting (2009–10), fishers have been required to report the number of net-lifts per fishing day, allowing CPUE (kg fisher day⁻¹) estimates to be standardised to the mean number of net-lifts per fisher-day.

The occurrence of ovigerous females in a large-scale, stratified, randomised survey of NSW identified their reproductive period to span from November to January (austral summer), peaking in December (Kennelly and Watkins 1994). In this survey, the estimated fecundity for individual crabs varied from ~78 000 to 256 000 eggs per batch and was positively correlated with body size

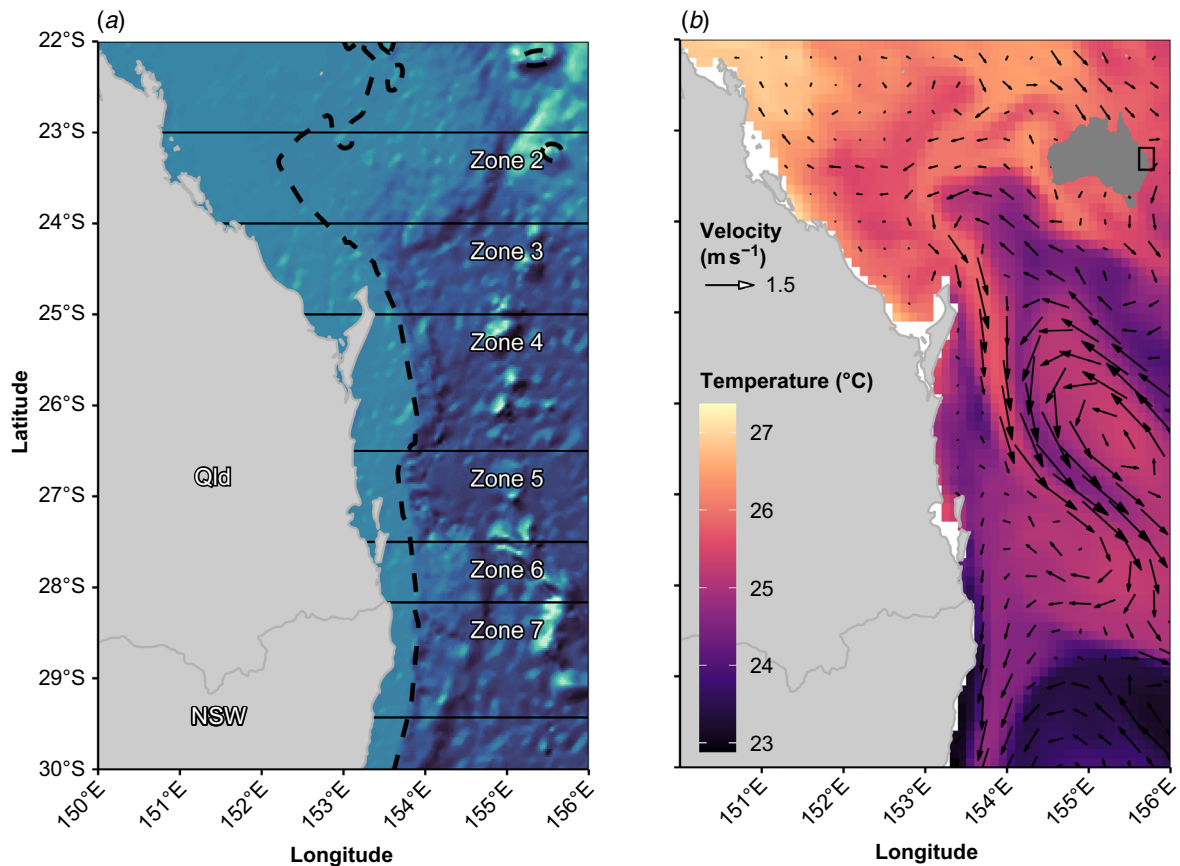


Fig. 1. Map of eastern Australia, showing (a) the six main fishing regions for spanner crabs overlaid on blended depth colours and relief shading of the ocean bottom derived from CleanTOPO2 data. The dashed line represents the 200-m isobath. Zones 2–6 are managed by Queensland (Qld), whereas Zone 7 is managed by New South Wales (NSW), and (b) an example (1 January 1993) of the oceanographic conditions common to this region as modelled in the BRAN2020 ocean model. The vectors show the direction and speed of the currents, and the colour represents the sea-surface temperature. For our simulations, particles were released on the benthos at depths of <200 m in all six zones shown in (a).

(Kennelly and Watkins 1994). Spanner crabs show limited movement following settlement and are estimated to enter the fishery at Age 4 (Brown 1986; Brown *et al.* 1999, 2008); however, age and growth remain uncertain for the species.

Hydrodynamic model

To simulate the dispersal of larval spanner crabs, velocity and temperature fields from the most recent (2020) version of the Bluelink ReANalysis (BRAN2020; Chamberlain *et al.* 2021) were used. BRAN2020 is a configuration of the Ocean Forecasting Australian Model, ver. 3 (OFAM3) and is an updated version of the previous BRAN models (Oke *et al.* 2013a, 2013b). A comprehensive description of OFAM3 is presented by Oke *et al.* (2013b). Briefly, OFAM3 is a near-global, eddy-resolving, z^* configuration of the Modular Ocean Model, developed principally for the purpose of reanalyses and forecasting upper ocean conditions worldwide. BRAN2020 takes observations from satellite sea-surface temperature (SST), satellite sea-level anomaly (SLA) and

in situ temperature and salinity and uses these to match the model output to real world conditions (full details in Chamberlain *et al.* 2021).

The analysis used the daily averaged outputs of zonal (east–west, u) and meridional (north–south, v) velocity (m s^{-1}) and temperature ($^{\circ}\text{C}$) within a subset of the model ($10\text{--}40^{\circ}\text{S}$, $140\text{--}165^{\circ}\text{E}$). This domain encompasses the entire eastern Australian region from the southern tip of Papua New Guinea (10°S) to the Bass Strait (40°S), stretching east to New Caledonia (165°E ; Fig. 1). Within this region the South Equatorial Current flows westward into the Coral Sea and Great Barrier Reef where it splits into equatorward (north) and poleward (south) components (Oke *et al.* 2019). The poleward component, the East Australian Current (EAC), is a western boundary current and is the dominant oceanographic feature along eastern Australia (Oke *et al.* 2019). The EAC forms a jet of fast-flowing water along the coastline, which generates upwelling and eddies along its path (Suthers *et al.* 2011; Cetina-Heredia *et al.* 2014; Schaeffer and Roughan 2015).

The model has a 0.1° (~ 10 -km) resolution grid, with a vertical resolution of 5 m in the upper ocean (top 40 m) and 10-m vertical resolution down to 200-m depth (Chamberlain *et al.* 2021). BRAN2020 is a data-assimilating model that continually adjusts the model to match observations, thereby recreating an accurate representation of past conditions, including mesoscale eddies with high accuracy (Chamberlain *et al.* 2021). The re-analysis model runs for 26 years (January 1993–December 2019).

Although BRAN2020 is fully validated and performs well at reconstructing mesoscale features (Chamberlain *et al.* 2021), an important but often overlooked part of using large-scale or global models in specific regions is the validation of the model against local, fine-scale observations. To assess the appropriateness and determine our confidence in the BRAN2020 model, we compared velocities and temperature at 10–30-m depth with observations at 13 ocean-observing moorings in our study region (Supplementary Table S1, Fig. S2). These moorings are independent and not assimilated into BRAN2020. We compared the modelled and observed velocities using Taylor diagrams to infer the relative strengths and weaknesses of BRAN2020 for our applications (Taylor 2001; North *et al.* 2009).

Particle tracking

Lagrangian particle-tracking simulations were conducted using PARCELS ('Probably A Really Computationally Efficient Lagrangian Simulator', ver. 2.2.0, see <https://oceanparcels.org/>; Lange and van Sebille 2017; Delandmeter and van Sebille 2019), run using Python (ver. 3.8.3, see <https://www.python.org/downloads/release/python-383/>; Van Rossum and Drake 2009). Lagrangian simulations were run using the horizontal velocities from all depth levels, with all vertical movement forced manually to simulate the biology of spanner crab larvae. During the spawning season, we released 1000 particles per day from each of the ocean zones (< 200 -m bathymetry). The location within each zone was randomised for each day. Because spawning is known to occur on the benthos and megalopa are found at the surface (Brown *et al.* 2008), all particles were released in the bottom ocean layer and then rose gradually to the surface over the next 10 days. As the swimming ability of spanner crab larvae is unknown, we have modelled all particles with passive horizontal dispersal (no swimming ability). Following a similar procedure to Schilling *et al.* (2020), the paths of each particle were interpolated using 5-min time steps, by using a fourth-order Runge–Kutta scheme, and recorded once daily. Each particle included a small random diffusion function to add natural variation to the movement of each particle. Following Okubo (1971), the scale of the horizontal diffusion constant ($8.8 \text{ m}^2 \text{ s}^{-1}$) was chosen to capture the effect of flows occurring at spatial scales smaller than that explicitly solved by the hydrodynamic model (e.g. sea breeze, turbulence). This parametrisation was successfully

used with a previous version of the Bluelink Reanalysis Model at the same resolution (Cetina-Heredia *et al.* 2015). Particles were, therefore, advected by both the BRAN2020 velocity fields and then in a random direction by the velocity scaled by the diffusion coefficient (van Sebille *et al.* 2018).

To account for the temperature-dependent nature of larval growth rates and pelagic larval durations, we modelled the 'settlement' of our larvae using degree-days (DD; Neuheimer and Taggart 2007; O'Connor *et al.* 2007). This approach calculates the cumulative daily temperatures experienced by larvae, and settlement is deemed to occur when the DD threshold is passed, resulting in faster settlement in warmer water. Experiments have shown that spanner crabs metamorphose to megalopa after 41.3 days at 25°C (Minagawa 1990). This is equivalent to 1032.5 DD, which was used as our threshold for settlement. As a sensitivity analysis, we tested multiple DD thresholds (800, 1000, 1200 and 1400 DD) and visually compared the predicted settlement patterns in our region of interest (NSW; Zone 7). The sensitivity analysis showed that DD thresholds of 200 (~ 7 –8 days) apart result in very similar settlement patterns, with differences more obvious with larger DD thresholds (Fig. S3). Regardless of the threshold tested, the broad patterns were consistent with all scenarios, showing putative peaks in settlement between 1997–2000 and 2012–2013. A longer threshold reduced the overall number of particles settling but because these are a relative measure, conclusions are unaffected.

We explicitly modelled natural mortality as part of our simulations by releasing many particles and applying a daily mortality rate to each cohort (particles released at the same time and place). Because natural mortality is assumed to be a daily constant for each cohort and each cohort started with the same number of particles, it can be assumed that the number of particles dying each day is equal among cohorts until particles begin to settle and the number of particles within each cohort may begin to vary (Schilling *et al.* 2020). Following this logic, to save computing time and reduce the required number of starting particles, we applied natural mortality only from Day 34 (prior to any particle reaching 1032.5 DD and settlement occurring). The actual mortality rate for spanner crab larvae in this region is unknown and we, therefore, applied the mortality rate from a laboratory study of spanner crab larvae, similar to Everett *et al.* (2017), which therefore does not include natural mortality from predation and is a conservative estimate of mortality. Regardless, by modelling a daily mortality rate (even if it is not entirely accurate) with a temperature-dependent growth rate, we are effectively able to model the effect of cumulative mortality on larval cohorts that have a longer PLD and are more vulnerable to predation before settlement. A simple sensitivity analysis showed that the predicted settlement under $10\times$ mortality was highly correlated with the original predicted settlement ($r = 0.96$,

$t_{22} = 17.31$, $P < 0.001$), demonstrating the outcomes in terms of relative settlement do not change in response to the specified mortality (because mortality is a constant). Larvae that did not reach the settlement threshold while on the continental shelf (<200-m bathymetry) within a recognised fishing zone were considered mortalities in terms of the analysis (dispersal mortality; Everett *et al.* 2017; Schilling *et al.* 2020). Owing to the course grid cells (~10 km) of BRAN2020, we assumed that settlement in <200-m bathymetry was representative of settlement in the 30–80-m bathymetric range.

Simulations and analysis

Assuming equal spawning output from each of the zones, we released 1000 particles per day in each zone during the spawning season (1 November–31 January) for 26 spawning seasons (1993–94 to 2018–19). Spawning seasons were identified as the year in which spawning began. On each day of the spawning season, a random location in each zone was selected and all particles were released there. This stratified random-release strategy ensures that over the spawning season there is a spread of spawning from shallower and deeper areas. All particles were then tracked for 60 days to ensure that they had sufficient time to reach the temperature-dependent settlement threshold.

Connectivity among the six fishing zones was assessed by two connectivity matrices, which can be viewed as a method to determine source–sink relationships within a population (Mitarai *et al.* 2009; North *et al.* 2009). For each spawning season, we calculated the percentage of larvae settling in each of the fishing zones (i.e. sinks) from each of the spawning regions (i.e. sources) as

$$PS_{ifh} = \frac{N_{ifh}}{T_{fh}} \times 100$$

where PS_{ifh} is the percentage settlement in Zone i , for particles released in Zone f in Spawning season h , N_{ifh} is the number of particles settling in Zone i , released in Zone f in Season h , T_{fh} is the total number of particles released in Zone f in Spawning season h .

We also calculated a second metric regarding the origins of particles landing within each zone as

$$PS_{ifh} = \frac{N_{ifh}}{G_{fh}} \times 100$$

where G_{fh} is the total number of particles landing in Zone f in Spawning season h . Averaging the percentage particles settled in each source–sink combination over the 26 spawning periods gives relative metrics (mean settlement) of connectivity among the different fishing areas. We also calculated the standard deviation of this mean, which provides information regarding the variability in the connectivity among the

zones. For each settlement zone, the first matrix shows the ‘origin percentages’ of particles settling in each spawning zone. The second matrix shows the ‘destination percentages’ of all spawned particles in each of the zones.

To test the hypothesis that strong recruitment events may be due to favourable oceanography within the NSW fishery, we quantified the number of particles predicted to settle in the NSW fishing zone (Zone 7) and assumed that this was a proxy for larval recruitment in each spawning season. This assumes equal spawning output each year across the stock, which is not realistic but without detailed information regarding the distribution and abundance of spawning crabs, this is a necessary assumption to examine our specific hypothesis. No detailed information exists regarding the distribution or interannual variation in the spawning stock biomass. We therefore expect that our metric at least partially captures overall recruitment, acknowledging that we are not accounting for variations in spawning stock biomass. Because the fishery comprises multiple year classes, we calculated a 3-year rolling average to our settlement predictions (to capture a main age class and the fast or slow growers from the following or preceding age classes). This was then lagged by 4 years, matching the broad assumption of the age at which male spanner crabs are estimated to recruit to the fishery (Brown 1986; Brown *et al.* 1999, 2008; Chen and Kennelly 1999). Because adult spanner crabs are non-migratory (Brown *et al.* 2008; Kapur *et al.* 2019), we consider this metric a good proxy for recruitment to the fishery, and suitable to assess how recruitment variability can contribute to variations in fisheries productivity. We, hereafter, refer to this metric as the ‘predicted recruitment index’.

To represent variation within the fishery, two CPUE metrics were calculated from NSW logbook data, both reported as the annual (July–June) average. The first source is a metric of kilograms of harvest per fisher-day, which extends from 1998–99 to 2019–20 (hereafter ‘long-term CPUE’). The second source is a metric of kilograms per net-lift and ranges from 2009–10 to 2019–20 (hereafter ‘short-term CPUE’) and which acknowledges that fishing effort (number of net lifts) may vary among days. The two CPUE metrics are highly correlated ($r = 0.82$, $t = 4.24$, $P = 0.002$). Within the NSW context, previous investigation showed no difference between using standardised catch rates or raw catch rates, so we used raw catch rates to maximise relevance to the fishery (Johnson 2022; detailed below). We used Pearson correlation tests to quantify how well our predicted recruitment index matched the observed fluctuations in each CPUE metric.

Standardised NSW catch rates predicted from generalised linear models (GLM) are presented in the NSW stock assessment for spanner crab (Johnson 2022). Explanatory model terms considered different catch rates among fishing years, seasons, individual fisher operations, their transformed fishing effort (the number of net-lifts, which was a function of

the number of ground-lines used, nets per ground-line and ground-line lifts per day; log or cube-root scale), the spatial locations of catches based on $6 \times 6'$ latitude and longitude grids. However, there is minimal difference between NSW raw and standardised catch rates (D. D. Johnson, unpubl. data), likely owing to stable management arrangements that were consistent, with no changes in permitted effort (capped at 30 dillies) or vessel capacity from 1997–98 to 2017–18 in NSW (<https://legislation.nsw.gov.au/view/html/inforce/current/sl-2006-0738#sch>).

In addition to the fishery-dependent CPUE, we also used a fishery-independent survey dataset that controls for vagaries of fisher behaviour, and has been identified as the best data source to investigate environmental influences on recruitment (Filar *et al.* 2021). Briefly, the survey is conducted annually each year (since 2005 in NSW), stratified by Fishing regions 2–7 (for full details of this survey, see Department of Primary Industries and Fisheries 2005; McGilvray *et al.* 2006). To focus again on NSW, we used data only from Region 7, which consisted of data from 2005 to 2018 inclusive. Data were extracted from Filar *et al.* (2021) by using the R package 'digitize' (ver. 0.0.4, see <https://CRAN.R-project.org/package=digitize>; Poisot 2011), and reported as the number of crabs per ground-line. The data were reported as sublegal crabs (<100-mm rostral carapace length), legal crabs (≥ 100 -mm rostral carapace length) and all crabs combined. The legal-size crab data were correlated with both fishery-dependent metrics (kg fisher-day⁻¹: $r = 0.84$, $P < 0.001$; kg net-lift⁻¹: $r = 0.64$, $P = 0.046$).

Spawning intensity is unlikely to be evenly distributed throughout the spawning season, and there is a clear peak during December (Kennelly and Watkins 1994). To investigate whether spawning in particular months was driving the relationship between our predicted recruitment index and CPUE, or whether spawning times may have changed during the past 25 years, we repeated the above analysis monthly from November to January. We hypothesised that the December relationship would be the strongest on the basis of the previously observed peak in spawning.

In our simulations, we assume that spawning is equal across space; however, this is unlikely to be the case. To assess the effect of violating this assumption, we conducted a sensitivity analysis where we sequentially removed spawning (i.e. released no particles) from within one region at a time. For each of these simulations, predicted (simulated) recruitment to the fishery in NSW was recalculated and compared with output from our full simulation (i.e. spawning in all regions) and other reduced simulations.

Oceanographic influences on connectivity

To investigate whether specific oceanographic features were contributing to years of high and low predicted settlement in terms of recruitment, we plotted the paths of particles that

successfully settled and particles that did not. By visually comparing the two groups of paths, we were able to make some broad conclusions of factors contributing to successful and unsuccessful settlement.

Results

Mooring comparison

Overall and meridional velocity from BRAN2020 was moderately correlated with observations at most of the moorings used in the comparison (Fig. S4). BRAN2020 performed well for the EAC and Coffs Harbour mooring, with a correlation generally between 0.4 and 0.5, but poorly for the shallower waters of the Great Barrier Reef moorings, with a correlation of < 0.3 , although the magnitudes of the velocities were quite similar (Fig. S4). The zonal velocity in BRAN2020 was a much poorer match to the observations, with mostly low correlations (< 0.2), except for the Coffs Harbour moorings, which were moderate (~ 0.4). Temperature was well resolved by BRAN2020 with all correlations > 0.7 and most > 0.9 . Overall, this suggests poor performance in the shallow areas with complex topography near the Great Barrier Reef to the north, and more reliable results in the EAC-influenced areas. The large-scale predominately along-shore patterns are well resolved (as already demonstrated by Chamberlain *et al.* 2021), whereas the small-scale patterns and cross-shelf transport are poorly resolved. BRAN2020 is therefore suitable for investigating broad-scale larval dispersal and connectivity (among zones), but is likely to not be suitable for small-scale movements such as within zones.

Connectivity

Larvae mostly settled south of their spawning location, although high levels of self-recruitment were evident within the northern zones. Zones 2 and 3 show relatively high amounts of self-recruitment and over 98% of larvae that settled in them originated within those two zones (Fig. 2a). By contrast, particles that settled further south originated from a larger area. Larvae settling in Zone 4 were primarily spawned in Zones 2, 3 and 4, whereas particles settling in Zones 5 through 7 were primarily spawned in Zones 3, 4 or 5 (Fig. 2a).

Taking a broader view and analysing the distribution of larvae as a proportion of total amount spawned in each zone, there is overall low survival (typical of natural mortality and dispersal), but a much higher settlement of particles spawned in the northern-most zones (2 and 3) than those spawned further south (Fig. 2b). Larvae spawned in Zones 4–7 were typically exported south from the inshore regions by the strong EAC, resulting in very low proportions of larvae from each of these zones settling because of a high dispersal mortality (Fig. 2b).

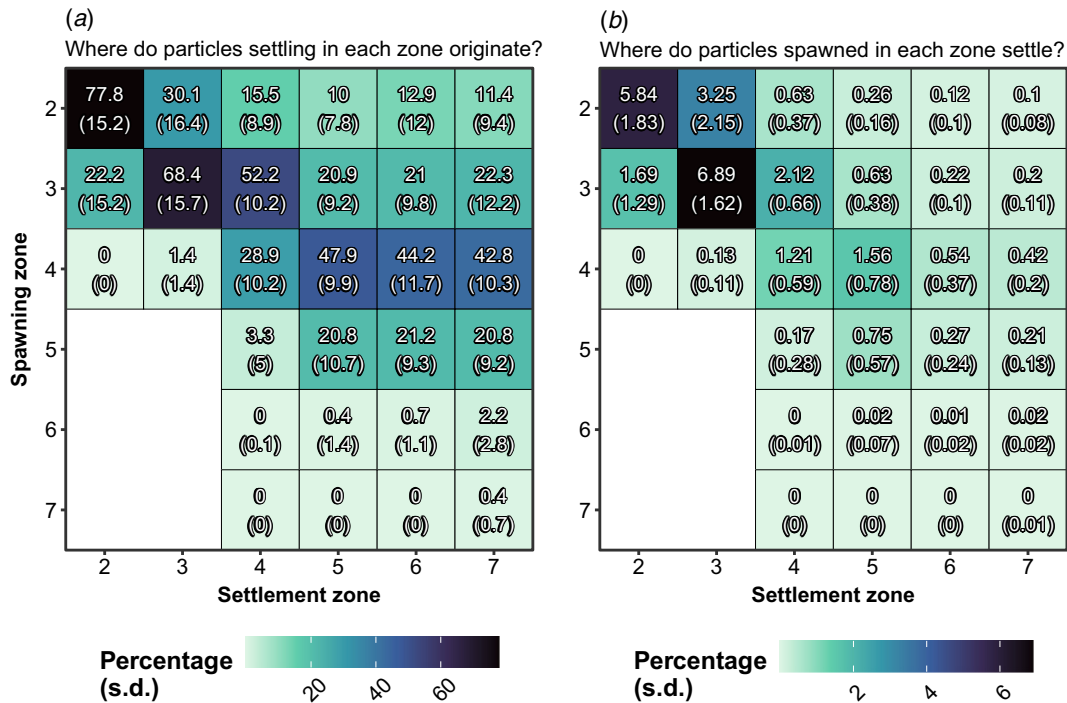


Fig. 2. Connectivity matrices based on the particle tracking simulations. (a) The origins of particles (y-axis; source) that settle in each zone (x-axis; sink). (b) The settlement locations (x-axis) of particles spawned in each zone (y-axis). Numbers show the average percentage over the model duration, with standard deviation shown in parentheses. The referred to zones can be seen in Fig. 1. Combinations of zones where connectivity was never observed are blank. Note that the different-magnitude colour scales differ among subplots.

Predicted settlement

The Qld region (Zones 2–6) received ~20× more larvae each year in our simulations than did NSW (Zone 7; Fig. 3). Within both regions, there were interannual variations; however, in terms of relative magnitude of the predicted settlement, the NSW region had larger fluctuations. In the NSW region, there was up to a 3× difference in predicted settlement among years. There were distinct periods of both high (1998–2000, 2008, 2012, 2013, 2015) and low (2001–2005, 2009–2011, 2014, 2018) settlement (Fig. 3).

Predicted recruitment to NSW fishery

There was evidence of a positive correlation between the predicted recruitment index (based on the 4-year lagged, rolling mean of settlement) and both CPUE time series as well as the fishery-independent survey data (Fig. 4, S5). There was evidence of a moderate correlation between the long-term CPUE and our predicted recruitment index ($r = 0.45$, $t_{20} = 2.262$, $P = 0.035$). Similarly, the short-term CPUE was positively correlated with the predicted recruitment index, but the evidence for the trend was weaker ($r = 0.48$, $t_9 = 1.623$, $P = 0.139$), which may be due to the timeseries being shorter (i.e. fewer observations and lower power). The strongest relationships were identified with the

legal-size crab data from the fishery-independent survey ($r = 0.73$, $t_{12} = 3.793$, $P = 0.003$). In the long-term series, there was an apparent breakdown in the correlations in the early 2000s, where predicted recruitment was high whereas CPUE was not.

The proportion of total harvest (Qld and NSW combined) taken in NSW showed strong evidence of being driven by recruitment. There was a moderate correlation with the predicted recruitment index ($r = 0.51$, $t_{21} = 2.262$, $P = 0.013$). When the predicted recruitment to NSW was higher, NSW made up a higher proportion of the total harvest. Although the relationships appear to correlate better in early years than in later years, the relationships generally correlated well with the exceptions of 2012 and 2013 (Fig. 5, S6).

Variation in settlement by spawning month and region

Overall patterns in settlement differed among months (Fig. 6). Settlement of November-spawned larvae was the most consistent, with high settlement predicted only in 1 high year (1993), contrasting with two peaks for December-spawned larvae (1998–2000, and 2012–2013). January-spawned larvae showed variation, but was overall more stable than was December-spawned larvae, with only one notable peak in predicted settlement in 2008.

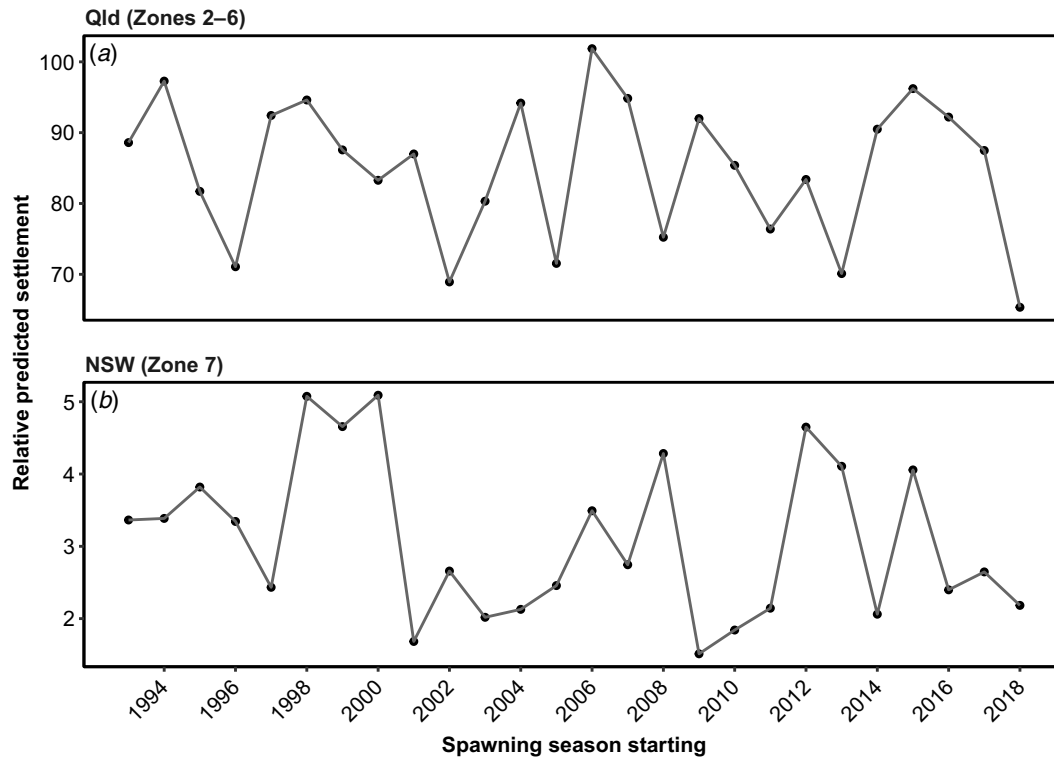


Fig. 3. Annual relative predicted settlement of spanner crab larvae in (a) Qld management region (Zones 2–6; Fig. 1) and, (b) the NSW management region (Zone 7; Fig. 1).

The pattern observed in the December-spawned larvae was noted to qualitatively match the observed patterns seen in both CPUE time series and fishery-independent survey data. After generating a predicted recruitment index to the fishery timeseries only on the basis of the December-spawned larvae, there was evidence of positive correlations with the CPUE time series (Fig. 7, S7). The long-term CPUE time series ($\text{kg fisher}\cdot\text{day}^{-1}$) showed strong evidence of a moderate correlation with the predicted recruitment index ($r = 0.49$, $t_{20} = 2.481$, $P = 0.022$). The shorter kilograms per net-lift time series showed a stronger positive correlation strength, but the evidence for the trend was only moderate because of the shorter time series ($r = 0.59$, $t_9 = 2.168$, $P = 0.059$). The fishery-independent survey data again showed the strongest relationships, with both all and legal-size crabs showing strong positive correlations (all crabs: $r = 0.78$, $t_{12} = 4.3378$, $P < 0.001$; legal-size crabs: $r = 0.84$, $t_{12} = 5.447$, $P < 0.001$). There was again a breakdown in the relationship between predicted recruitment index and CPUE in the early 2000s, seen in the longer time series.

The sensitivity analysis testing the removal of spawning from each region showed that removing spawning entirely from any of the regions did not significantly alter the predicted recruitment to the NSW fishery. All variations were highly correlated with the equal spawning assumption ($r \geq 0.92$) and showed the same peaks and troughs

(Fig. S8), suggesting that the reported patterns are robust to spatial variations in spawning.

Successful transport drivers

When comparing the paths of particles that successfully settled and those that did not, it was evident that excess southward transport was detrimental, as no particles returned north and successfully settled if they went south of 30°S (Fig. 8). At $\sim 25^\circ\text{S}$, Fraser Island (K'gari) extends into the Pacific Ocean, forming a 'choke point' where many larvae were retained to the north-west of the island, driving the high settlement rates in this zone (Zone 3), with very few of the particles being retained by the island, then leaving (Fig. 8a). In the northern fishing zones, there was a large persistent eddy that enabled some particles in $23\text{--}24^\circ\text{S}$ to be transported off the shelf and still be returned (Fig. 8b). Further south, there were decreasing numbers of tracks that successfully returned to the shelf after being transported into deeper waters (Fig. 8b).

Discussion

This paper has demonstrated that oceanographic drivers are likely to influence dispersal and connectivity across the Australian spanner crab stock. The model suggests that the

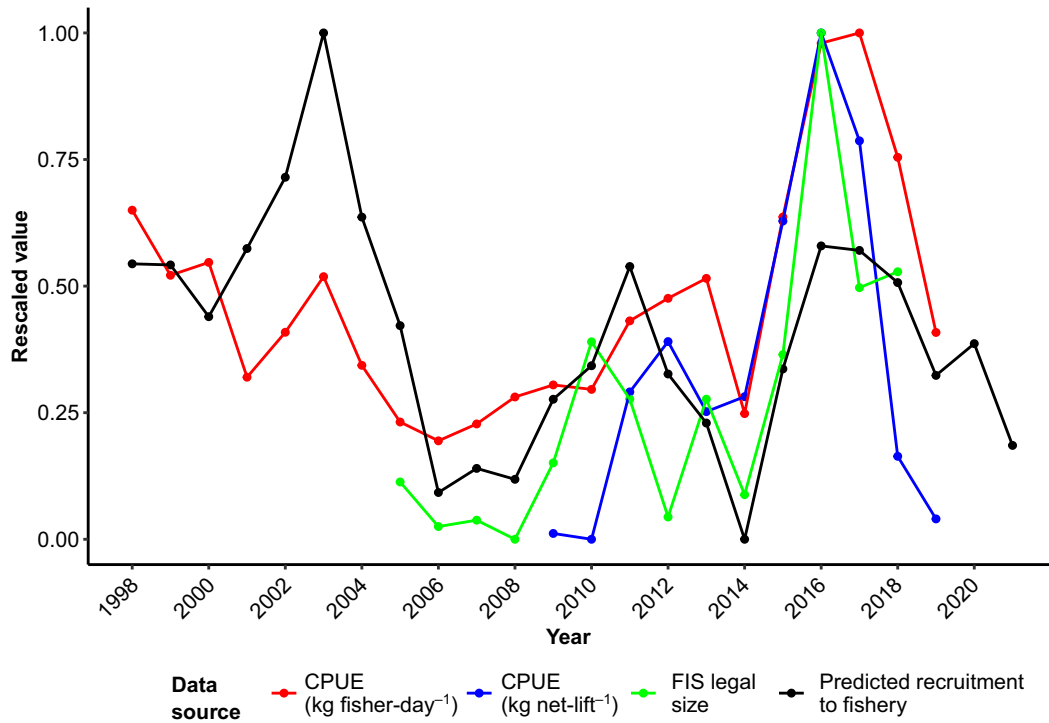


Fig. 4. Relationship between the NSW CPUE metrics and the predicted recruitment to the spanner crab fishery based on the larval settlement estimates. The black line shows the predicted recruitment to the fishery based on a 3-year rolling mean of predicted larval settlement and a 4-year lag. The red line shows the CPUE based on kilograms of harvest per fisher-day. The blue line shows the CPUE based on kilograms per net-lift. The green line shows the legal-size crab CPUE from the fishery-independent survey. Note that the y-axis was rescaled to assist interpretation.

eastern Australian spanner crab stock is highly likely to be supported by larvae spawned within the northern half of the available fishing area, within the Qld jurisdiction. Modelling also indicated that larvae spawned within the southern portions of the Qld fishery and NSW are likely to be transported away from the suitable habitat areas (<200-m bathymetry within the fishing zones) by the East Australian Current, limiting the potential contribution of the southern zones to recruitment. Variation in CPUE in the southern fishing zone (both fishery-dependent and fishery-independent) correlated with predicted recruitment to the fishery solely on the basis of simulated larval dispersal. This indicated that fluctuations in the NSW spanner crab fishery are likely to be recruitment driven and that there may be potential to forecast abundance. Further, predicted recruitment can be used as a performance indicator for the stock and may support the development of decision rules within a harvest strategy.

Connectivity

There were distinct patterns in connectivity among spawning and settlement locations for spanner crabs. The two northern zones (Zones 2 and 3) showed high levels of self-recruitment,

suggesting that these zones may be self-sustaining in terms of long-term larval supply. This contrasted with zones further south, with the probability of self-recruitment decreasing with increasing latitude. This is likely to arise because of the prevailing oceanography of the region, particularly the increasing influence of the EAC because its higher proximity to the southern spawning areas increases the likelihood of transporting the larvae south, away from possible settlement zones. The general observation for the southern zones is that almost all the larval supply comes from the north, transported by the EAC, an observation consistent with previous larval dispersal studies south of the current study (Roughan *et al.* 2011; Everett *et al.* 2017; Schilling *et al.* 2020).

A similar latitudinal gradient was seen in the differential 'survival' from the different zones. Successful settlement was an order of magnitude higher in the northern zones than in the southern zones, with successful settlement dropping from >5% of all spawned larvae in Zones 2 and 3 to <4% in Zone 4 and <1% in Zones 6 and 7. This is again likely to be due to the influence of the EAC, in addition to the fact that the area available for settlement in the southern zones is substantially smaller than in the northern zones, which have a wider continental shelf. The extension of Fraser

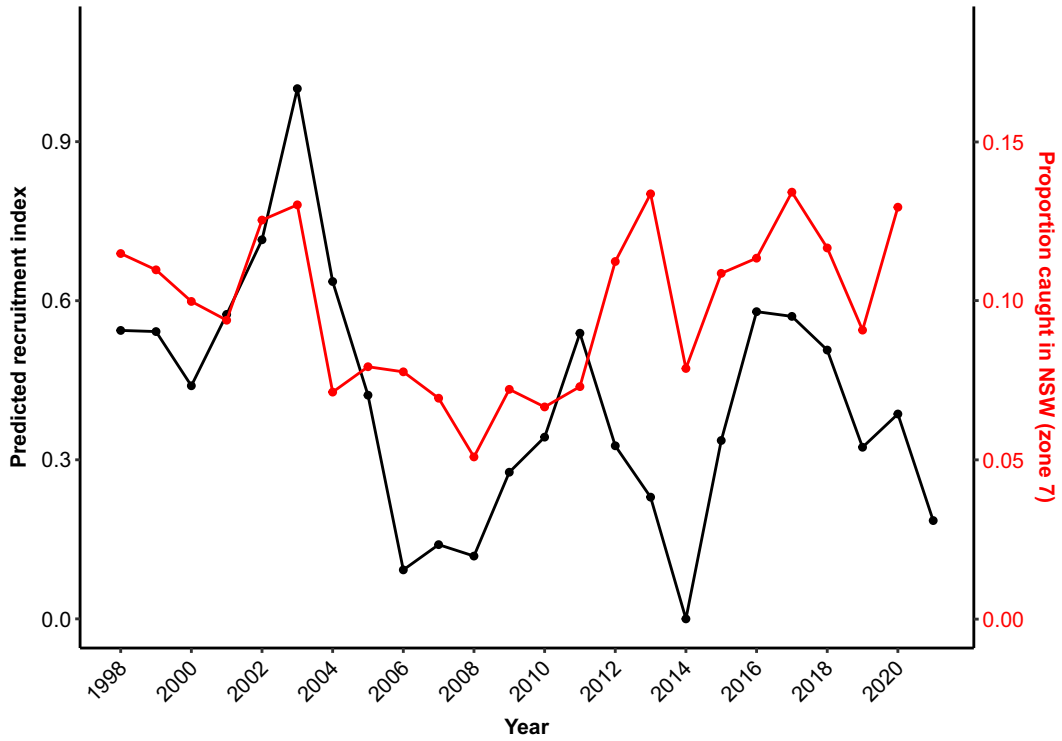


Fig. 5. Relationship between the predicted recruitment index (black) and the proportion of total harvest taken in NSW (red; Zone 7, Fig. 1). The predicted recruitment index is based on a 3-year rolling mean of predicted larval settlement with a 4-year lag.

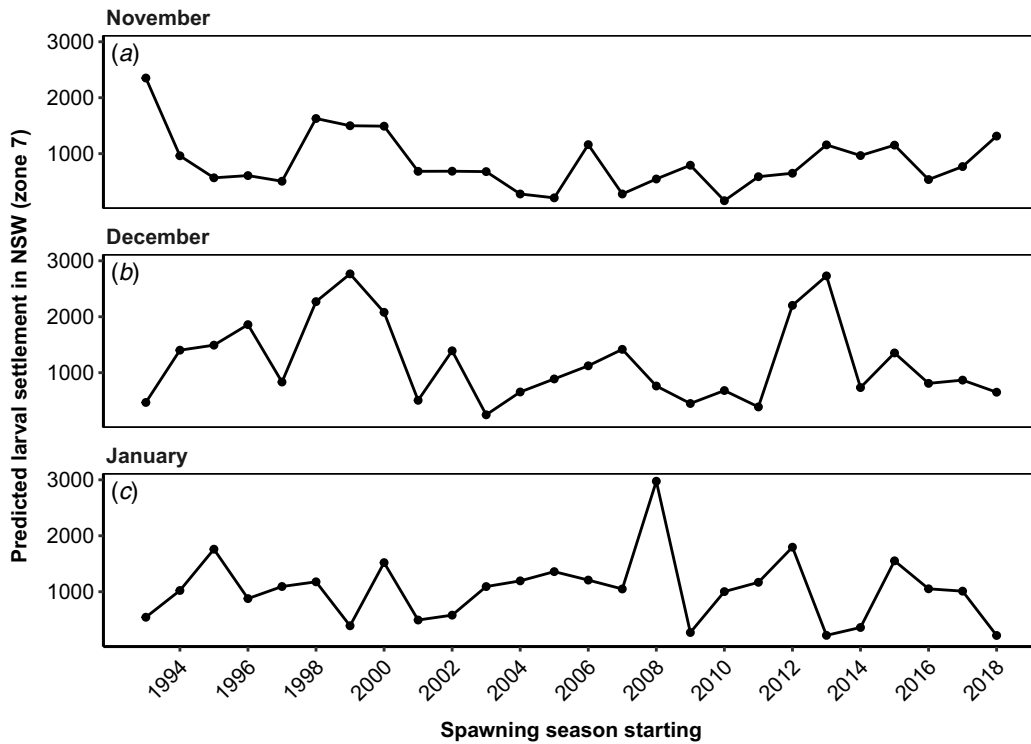


Fig. 6. Predicted settlement in NSW (Zone 7, Fig. 1) for larvae spawned in each month of our simulated spawning season: (a) November; (b) December; and (c) January. The y-axis is comparable among subplots, but is a relative number.

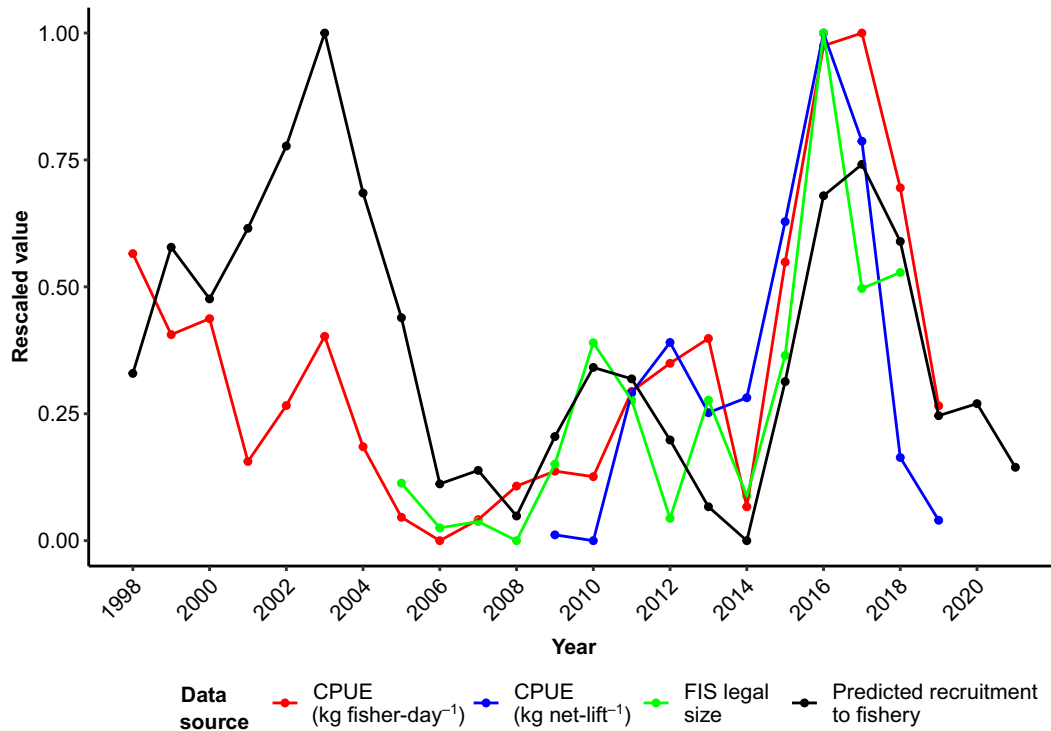


Fig. 7. Correlation between the two NSW CPUE metrics and the predicted recruitment to the spanner crab fishery based on December-spawned larvae settlement estimates. The black line shows the predicted recruitment index based on a 3-year rolling mean of predicted larval settlement and a 4-year lag. The red line shows the CPUE based on kilograms of harvest per fisher-day. The blue line shows the CPUE based on kilograms per net lift. The green line shows the relative CPUE for legal-size crabs from the fishery-independent survey. Note that the y-axis was rescaled to assist interpretation.

Island out into the Pacific Ocean was also a significant feature that increased settlement by capturing larvae inshore, out of the EAC. The eastern side of Fraser Island is recognised as where the EAC is most coherent (Kerry *et al.* 2018), and larvae in this region will be transported southward rapidly. It is interesting to note that much of the stock is now located in the regions around Fraser Island, potentially contributing to the high recruitment in NSW via the dominant southward current.

Recruitment within NSW

We found moderate to strong correlations between the predicted recruitment and the observed variation in catch rates, depending on the metric used. The strongest relationship was found using December-simulated spawning and the legal-size crabs in the fishery-independent survey, aligning with the current belief that November–December is the peak spawning period (Skinner and Hill 1987; Kennelly and Watkins 1994). The predicted recruitment index from the larval dispersal simulations closely matched NSW CPUE in recent years, but the relationship broke down in the early 2000s. We believe this could be the effect of varying levels of spawning stock biomass resulting in an

incorrect number of particles released in those years (e.g. perhaps the overall numbers spawned should be higher or lower in some years), although it is recognised that it is likely that abundances within the Qld fishery declined from the early 2000s to 2017 (Roelofs *et al.* 2021). A more likely alternative may be a shift in the dominant spawning biomass from Region 3 to Region 4, aligning with shifts in both the harvest and fishery-independent survey, which occurred around this time (Filar *et al.* 2021). This dynamic would not have been captured in our assumption of equal spawning in time and space; however, our sensitivity analysis showed that the overall patterns in the results would not have been altered.

Management and assessment implications

When the settlement and survival patterns are viewed together, it strongly suggests that both the Qld (Zones 2–6) and NSW (Zone 7) zones are supplied by larvae from the Qld zones. As such, changes in larval supply from Qld (arising through a reduction in egg and larvae production or changes in oceanographic features that affect larval transport) may pose a risk to both NSW and Qld fisheries. The limited survival of larvae spawned within the NSW reporting zone

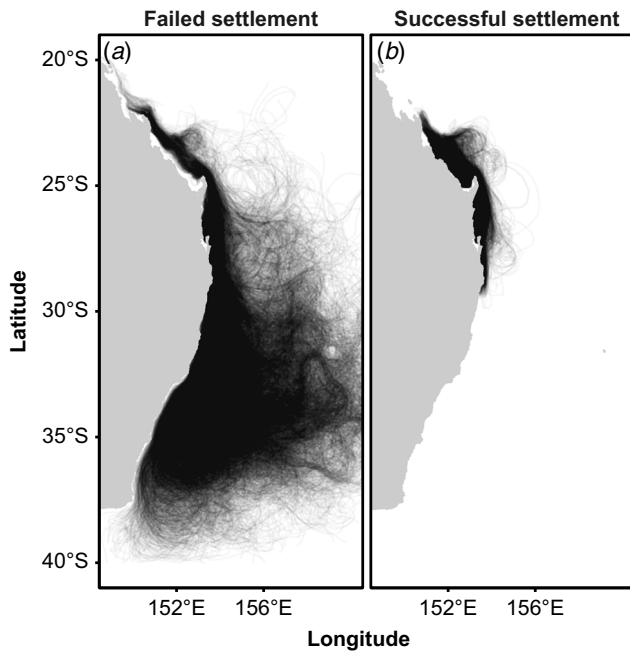


Fig. 8. A random subset of tracks from larvae that (a) failed to settle in an appropriate area and, (b) successfully settled in an appropriate area. Successful settlement was defined as reaching the degree days threshold while on the continental shelf and within the fishing zones (2–7, Fig. 1).

suggests that the NSW fishery is heavily subsidised by the Qld spawning stock.

The dependency of the entire stock on larval supply from the Qld jurisdiction highlights the need to maintain a healthy spawning stock in this region. Any decline in spawning stock biomass and subsequent larval supply from Qld region, in particular from Zones 3–5, could have implications across the whole fishery. Going forward, the strong connectivity results here suggest that it may be wise to combine Qld and NSW management and assessment procedures.

A detailed study regarding the reproductive biology, including fecundity over the latitudinal range, would be advantageous to advance these particle-tracking modelling scenarios further beyond the current study. Recent research has also shown 4-year lagged effects of both chlorophyll-*a* (positive effect) and sea-level anomaly (negative effect) on catch rates in Region 3 (Filar *et al.* 2021); this suggests that there could be important influences beyond larval supply, such as the influence of food availability for larvae (Leggett and Deblois 1994), or mesoscale oceanographic effects such as eddies (Bakun 2006). Management of the Qld fishery is currently undertaken with a range of input and output controls, with the catch being limited by a total allowable catch (TAC). A target spawning biomass ratio of 60% has been imposed as part of a harvest strategy based on a maximum economic yield (Queensland Government 2019).

If achieved, this 60% target should provide some security for both Qld and NSW fisheries in terms of continued larval supply.

An increasingly important factor in eastern Australia is climate change. Both northern and southern eastern Australia are being rapidly affected by warming waters and shifting currents. The Great Barrier Reef to the north is experiencing sustained heatwaves, resulting in increased bleaching and impaired coral recruitment (Hughes *et al.* 2017, 2019), occurring concurrently with a strengthening EAC (Sun *et al.* 2012; Li *et al.* 2021), resulting in increased water temperatures at higher latitudes, creating a global warming hotspot (Hobday and Pecl 2014). Located between these dramatic changes, the spanner crab habitat is warming, which may lead to a southward redistribution through time if the north of the range is unfavourably warm (Pecl *et al.* 2017; Gervais *et al.* 2021), although it is unknown whether the habitat further south would be suitable. It is likely that a climate-driven shift may have already occurred, with the catch in Region 3 declining, resulting in Region 4 becoming the most productive (Filar *et al.* 2021). If such a southward shift continues, larval supply could be detrimentally affected, with an increased southerly transport.

The interannual variations in predicted settlement in both Qld and NSW have implications for future CPUE-based stock assessments, and potential limitations on predicted biomass that is available for harvest. For example, models that operate on the assumption of average recruitment across all stock sizes, including Surplus Production Models (SPMs) previously applied to the Hawaiian spanner crab fishery (Kapur *et al.* 2019), are sensitive to potential changes and the lagged effects of recruitment (Punt and Szuwalski 2012). On the basis of predicted recruitment, estimates of maximum sustainable yield (MSY) and associated fisheries reference points derived from fitting SPMs to data from jurisdictions separately without adjustments based on predicted recruitment dynamics, may be biased.

Limitations

Although our simulations showed strong agreement to patterns of catch, there are several factors potentially contributing to the disparity between predicted recruitment and fisheries harvest in some years. Some of these factors include (1) the 4-year time-lag between recruitment and harvest, (2) our estimation of larval mortality, (3) the disparity between the hydrodynamic model and reality, (4) the exclusion of swimming for our larvae, and (5) the uncertainty surrounding the biology of spanner crabs. Our predicted recruitment to the fishery was based on estimates that male spanner crabs enter the fishery at 4 years old (Brown *et al.* 2008). This is an assumption that carries considerable uncertainty, with previous minimum legal-size estimates ranging from 1.75 to 8.83 years for females and from 1.08 to 3.58 years for males (Kirkwood *et al.* 2005).

It would be extremely valuable for further studies to better define and understand growth rates as well as the settlement of post-larvae (e.g. where 1–3-year-old juveniles primarily reside). In these 4 years, it is probable that other factors, such as environmental fluctuations, influenced the number of juveniles reaching maturity. Environmental factors such as temperature and current speed have been shown to influence the catchability of spanner crabs in this region (Spencer *et al.* 2017, 2019a, 2019b, 2019c), and it is possible that such factors may also influence juvenile survival. Environmental variables such as onshore winds have also shown correlations with larval fish supply in this region (Schilling *et al.* 2022). Swimming ability has also previously been demonstrated to significantly improve dispersal modelling of juvenile and larval fish, and turtles (Putman and Mansfield 2015; Leis 2021). Unfortunately, there is scant information regarding the swimming ability (speed or direction) of spanner crab larvae and it was, therefore, not possible to include it in our models. As the adult habitat of spanner crab larvae is the continental shelf of Australia, we suggest that swimming would generally occur in a westward direction and the overall connectivity patterns would remain similar because of the predominant flow of the currents along the coast, although overall recruitment might be higher.

There is considerable uncertainty around the reproductive biology of spanner crabs because of most of the reproductive information being over 20 years old. Predicted recruitment was simulated from different months as a control to test whether there have been possible shifts in spawning season. The collection of new data to describe the size–fecundity relationship at smaller spatial scales than previously investigated, and ongoing monitoring of the timing and duration of the spawning period in the Qld fishery, may provide a more accurate representation of the actual egg production to refine future simulations. Replicating existing annual standardised fishery-independent surveys during the identified spawning period would provide direct estimates of the abundance, size structure and the proportion of female crabs spawning. This would also allow for the collection of data to monitor spatial and temporal variations in the distribution and abundance of the spawning population, possibly even with reference to optimal sea-surface temperatures for spawning. This is particularly important, with rapid ocean warming occurring within the EAC region (Wu *et al.* 2012), and temperature clearly affecting fecundity in other decapods (Johnston and Yeoh 2021; Nolan *et al.* 2022). However, the cost–benefit of increased fishery-independent monitoring and the relative value of data on spawning populations (e.g. cost–benefit ratio on TAC) within an integrated fishery model for the East Coast biological stock would need to be determined to support this recommendation (Dennis *et al.* 2015).

Our assessment of the BRAN2020 model identified that it performed well in terms of temperature reconstruction and

has previously been demonstrated to accurately represent large-scale dynamics (Chamberlain *et al.* 2021). However, BRAN2020 did not resolve small-scale dynamics, particularly cross-shelf velocities, adequately (Fig. S4). This may have resulted in the fine-scale movement dynamics of our larvae being poorly represented. To overcome this, it would be necessary to use a higher-resolution ocean model that better represents fine-scale dynamics on the continental shelf. Such models are available, but currently do not fully cover the area of interest with the main areas of interest in the models, either focusing north on the Great Barrier Reef (Steven *et al.* 2019; ~9–28.5°S) or south on the Separation zone of the EAC (Kerry *et al.* 2018; ~26–40°S). These models would need to have an extended domain or be nested within each other; both options are not currently available. Regardless, although the lack of fine-scale representation may mean that within-zone movement is poorly represented, we are confident that the large-scale connectivity findings are robust, with transport dominated by southerly currents. The strong correlations between lagged predicted recruitment and observed fluctuations in CPUE provide us with confidence in the general ability of our model and suggest that the dispersal is most influenced by larger mesoscale ocean dynamics.

Conclusions

The evidence presented here suggests that recruitment within the spanner crab stock in eastern Australia is likely to be heavily supported by spawning in the northern half of the fishing area. Larvae spawned from the northern and middle zones showed higher rates of successful settlement in all fishing areas than did larvae spawned in the more southern zones where the EAC has strong influence. In the NSW managed zone (Zone 7), we showed significant correlations between predicted recruitment to the fishery (using commercial and independent CPUE metrics) and modelled larval settlement. Combined with the connectivity results, this suggests that the NSW fishery is supported by recruits originating from the Qld fishery (Zones 2–6) and that joint management and assessments between the jurisdictions may be beneficial. By monitoring the arrival of larval or juvenile spanner crabs in NSW and using improved ocean models, it may be possible to reliably forecast future years of anomalously high and low CPUE in the NSW fishery.

Supplementary material

Supplementary material is available [online](#).

References

- Anderson CNK, Hsieh C, Sandin SA, Hewitt R, Hollowed A, Beddington J, May RM, Sugihara G (2008) Why fishing magnifies fluctuations in fish abundance. *Nature* 452, 835–839. doi:10.1038/nature06851

- Bakun A (2006) Fronts and eddies as key structures in the habitat of marine fish larvae: opportunity, adaptive response and competitive advantage. *Scientia Marina* **70**, 105–122. doi:10.3989/scimar.2006.70s2105
- Brown IW (1986) Population biology of the spanner crab (*Ranina ranina*) in south-east Queensland. Unpublish Report Queensland Department of Primary Industries, Brisbane, Qld, Australia.
- Brown IW, Kirkwood J, Gaddes S, Dichmont C, Ovenden J (1999) Population dynamics and management of spanner crabs (*Ranina ranina*) in southern Queensland. Final report on FRDC Project 95/022, Department of Primary Industries.
- Brown IW, Scandol J, Mayer DG, Campbell MJ, Kondylas S, McLennan SR, Williams AJ, Krusic-Golub K, Treloar T (2008) Reducing uncertainty in the assessment of the Australian spanner crab fishery. Final report on FRDF Project 2003/046, Fisheries Research and Development Corporation, Canberra, ACT, Australia.
- Cetina-Heredia P, Roughan M, Liggins G, Coleman MA, Jeffs A (2019) Mesoscale circulation determines broad spatio-temporal settlement patterns of lobster. *PLoS One* **14**, e0211722. doi:10.1371/journal.pone.0211722
- Cetina-Heredia P, Roughan M, van Sebille E, Coleman MA (2014) Long-term trends in the East Australian Current separation latitude and eddy driven transport. *Journal of Geophysical Research: Oceans* **119**, 4351–4366. doi:10.1002/2014JC010071
- Cetina-Heredia P, Roughan M, van Sebille E, Feng M, Coleman MA (2015) Strengthened currents override the effect of warming on lobster larval dispersal and survival. *Global Change Biology* **21**, 4377–4386. doi:10.1111/gcb.13063
- Chamberlain MA, Oke PR, Fiedler RAS, Beggs HM, Brassington GB, Divakaran P (2021) Next generation of BlueLink ocean reanalysis with multiscale data assimilation: BRAN2020. *Earth System Science Data* **13**, 5663–5688. doi:10.5194/essd-13-5663-2021
- Chen Y, Kennelly SJ (1999) Growth of spanner crabs, *Ranina ranina*, off the east coast of Australia. *Marine and Freshwater Research* **50**, 319–325. doi:10.1071/MF98023
- Coleman MA, Feng M, Roughan M, Cetina-Heredia P, Connell SD (2013) Temperate shelf water dispersal by Australian boundary currents: implications for population connectivity. *Limnology and Oceanography: Fluids and Environments* **3**, 295–309. doi:10.1215/21573689-2409306
- Cowen RK, Gawarkiewicz G, Pineda J, Thorrold SR, Werner FE (2007) Population connectivity in marine systems an overview. *Oceanography* **20**, 14–21. doi:10.5670/oceanog.2007.26
- Delandmeter P, van Sebille E (2019) The Parcels v2.0 Lagrangian framework: new field interpolation schemes. *Geoscientific Model Development* **12**, 3571–3584. doi:10.5194/gmd-12-3571-2019
- Dennis D, Plagányi É, Van Putten I, Hutton T, Pascoe S (2015) Cost benefit of fishery-independent surveys: are they worth the money? *Marine Policy* **58**, 108–115. doi:10.1016/j.marpol.2015.04.016
- Department of Primary Industries and Fisheries (2005) Fisheries long term monitoring program sampling protocol – spanner crab: (2000 onwards) section 1. (Department of Primary Industries and Fisheries: Brisbane, Qld, Australia) Available at https://www.daf.qld.gov.au/_data/assets/pdf_file/0004/57514/SpannerCrab-Sampling.pdf
- Dichmont CM, Brown IW (2010) A case study in successful management of a data-poor fishery using simple decision rules: the Queensland spanner crab fishery. *Marine and Coastal Fisheries* **2**, 1–13. doi:10.1577/C08-034.1
- Dixon PA, Milicich MJ, Sugihara G (1999) Episodic fluctuations in larval supply. *Science* **283**, 1528–1530. doi:10.1126/science.283.5407.1528
- Doblin MA, van Sebille E (2016) Drift in ocean currents impacts intergenerational microbial exposure to temperature. *Proceedings of the National Academy of Sciences of the United States of America* **113**, 5700–5705. doi:10.1073/pnas.1521093113
- Everett JD, van Sebille E, Taylor MD, Suthers IM, Setio C, Cetina-Heredia P, Smith JA (2017) Dispersal of eastern king prawn larvae in a western boundary current: new insights from particle tracking. *Fisheries Oceanography* **26**, 513–525. doi:10.1111/fog.12213
- Filar JA, Courtney AJ, Gibson LJ, Jemison R, Leahy SM, Lei Y, Mendiolar M, Mitchell J, Robson B, Steinberg C (2021) Modelling environmental changes and effects on wild-caught species in Queensland. Department of Agriculture and Fisheries, Queensland, The University of Queensland and the Australian Institute of Marine Science, Brisbane, Qld, Australia.
- Gallego A, North EW, Petitgas P (2007) Introduction: status and future of modelling physical–biological interactions during the early life of fishes. *Marine Ecology Progress Series* **347**, 121–126. doi:10.3354/meps06972
- Gervais CR, Champion C, Pecl GT (2021) Species on the move around the Australian coastline: a continental-scale review of climate-driven species redistribution in marine systems. *Global Change Biology* **27**, 3200–3217. doi:10.1111/gcb.15634
- Hinrichsen H-H, Dickey-Collas M, Huret M, Peck MA, Vikebø FB (2011) Evaluating the suitability of coupled biophysical models for fishery management. *ICES Journal of Marine Science* **68**, 1478–1487. doi:10.1093/icesjms/fsr056
- Hobday AJ, Pecl GT (2014) Identification of global marine hotspots: sentinels for change and vanguards for adaptation action. *Reviews in Fish Biology and Fisheries* **24**, 415–425. doi:10.1007/s11160-013-9326-6
- Hughes TP, Kerry JT, Álvarez-Noriega M, Álvarez-Romero JG, Anderson KD, Baird AH, Babcock RC, Beger M, Bellwood DR, Berkemans R, Bridge TC, Butler IR, Byrne M, Cantin NE, Comeau S, Connolly SR, Cumming GS, Dalton SJ, Diaz-Pulido G, Eakin CM, Figueira WF, Gilmour JP, Harrison HB, Heron SF, Hoey AS, Hobbs J-PA, Hoogenboom MO, Kennedy EV, Kuo C-y, Lough JM, Lowe RJ, Liu G, McCulloch MT, Malcolm HA, McWilliam MJ, Pandolfi JM, Pears RJ, Pratchett MS, Schoepf V, Simpson T, Skirving WJ, Sommer B, Torda G, Wachenfeld DR, Willis BL, Wilson SK (2017) Global warming and recurrent mass bleaching of corals. *Nature* **543**, 373–377. doi:10.1038/nature21707
- Hughes TP, Kerry JT, Baird AH, Connolly SR, Chase TJ, Dietzel A, Hill T, Hoey AS, Hoogenboom MO, Jacobson M, Kerswell A, Madin JS, Mieog A, Paley AS, Pratchett MS, Torda G, Woods RM (2019) Global warming impairs stock-recruitment dynamics of corals. *Nature* **568**, 387–390. doi:10.1038/s41586-019-1081-y
- Johnson D (2021) NSW stock status summary 2020/21 – spanner crab – (*Ranina ranina*). NSW Department of Primary Industries, Fisheries NSW, Australia.
- Johnson D (2022) NSW stock status summary 2021/22 – spanner crab – (*Ranina ranina*). (NSW Department of Primary Industries, Fisheries NSW, Australia) Available at https://www.dpi.nsw.gov.au/_data/assets/pdf_file/0008/1392038/stock-status-summary-2021-22-spanner-crab-2.pdf
- Johnston DJ, Yeoh DE (2021) Temperature drives spatial and temporal variation in the reproductive biology of the blue swimmer crab *Portunus armatus* A. Milne-Edwards, 1861 (Decapoda: Brachyura: Portunidae). *Journal of Crustacean Biology* **41**, ruab032. doi:10.1093/jcbl/ruab032
- Kapur M, Fitchett MD, Yau A, Carvalho F (2019) 2018 benchmark stock assessment of main Hawaiian islands kona crab. NOAA Technical Memorandum NMFS-PIFSC-77, National Oceanic and Atmospheric Administration, National Marine Fisheries Service, Pacific Islands Fisheries Science Center. doi:10.25923/7wf2-f040
- Kennelly SJ, Watkins D (1994) Fecundity and reproductive period, and their relationship to catch rates of spanner crabs, *Ranina ranina*, off the east coast of Australia. *Journal of Crustacean Biology* **14**, 146–150. doi:10.2307/1549061
- Kerry C, Roughan M, Powell B (2018) Observation impact in a regional reanalysis of the East Australian Current system. *Journal of Geophysical Research: Oceans* **123**, 7511–7528. doi:10.1029/2017JC013685
- Kirkwood JM, Brown IW, Gaddes SW, Hoyle S (2005) Juvenile length-at-age data reveal that spanner crabs (*Ranina ranina*) grow slowly. *Marine Biology* **147**, 331–339. doi:10.1007/s00227-005-1574-0
- Kurota H, Szuwalski CS, Ichinokawa M (2020) Drivers of recruitment dynamics in Japanese major fisheries resources: effects of environmental conditions and spawner abundance. *Fisheries Research* **221**, 105353. doi:10.1016/j.fishres.2019.105353
- Lange M, van Sebille E (2017) Parcels v0.9: prototyping a Lagrangian ocean analysis framework for the petascale age. *Geoscientific Model Development* **10**, 4175–4186. doi:10.5194/gmd-10-4175-2017
- Leggett WC, Deblois E (1994) Recruitment in marine fishes: is it regulated by starvation and predation in the egg and larval stages? *Netherlands*

- Journal of Sea Research* **32**, 119–134. doi:10.1016/0077-7579(94)90036-1
- Leis JM (2021) Perspectives on larval behaviour in biophysical modelling of larval dispersal in marine, demersal fishes. *Oceans* **2**, 1–25. doi:10.3390/oceans2010001
- Li J, Roughan M, Kerry C (2021) Dynamics of interannual eddy kinetic energy modulations in a Western Boundary Current. *Geophysical Research Letters* **48**, e2021GL094115. doi:10.1029/2021gl094115
- Lowe WH, Allendorf FW (2010) What can genetics tell us about population connectivity? *Molecular Ecology* **19**, 3038–3051. doi:10.1111/j.1365-294X.2010.04688.x
- McGilvray J, Brown I, Smallwood D (2006) Fisheries long term monitoring program – summary of spanner crab (*Ranina ranina*) survey results: 2000–2005. (Department of Primary Industries and Fisheries: Brisbane, Qld, Australia) Available at <https://era.daf.qld.gov.au/id/eprint/6423/1/Spanner-Report-2000-2005.pdf>
- Minagawa M (1990) Influence of temperature on survival, feeding and development of larvae of the red frog crab, *Ranina ranina* (Crustacea, Decapoda, Raninidae). *Nippon Suisan Gakkaishi* **56**, 755–760. doi:10.2331/suisan.56.755
- Mitarai S, Siegel DA, Watson JR, Dong C, McWilliams JC (2009) Quantifying connectivity in the coastal ocean with application to the Southern California Bight. *Journal of Geophysical Research: Oceans* **114**, C10026. doi:10.1029/2008jc005166
- Neuheimer AB, Taggart CT (2007) The growing degree-day and fish size-at-age: the overlooked metric. *Canadian Journal of Fisheries and Aquatic Sciences* **64**, 375–385. doi:10.1139/f07-003
- Nolan SEF, Johnson DD, Hanamseth R, Suthers IM, Taylor MD (2022) Reproductive biology of female blue swimmer crabs in the temperate estuaries of south-eastern Australia. *Marine and Freshwater Research* **73**, 366–376. doi:10.1071/MF21191
- Norrie C, Dunphy B, Roughan M, Weppe S, Lundquist C (2020) Spill-over from aquaculture may provide a larval subsidy for the restoration of mussel reefs. *Aquaculture Environment Interactions* **12**, 231–249. doi:10.3354/aei00363
- North EW, Gallego A, Petitgas P (2009) Manual of recommended practices for modelling physical–biological interactions during fish early life. ICES Cooperative Research Report 295. International Council for the Exploration of the Sea.
- NSW Total Allowable Fishing Committee (2020) Report and determinations for the 2020–21 fishing period nsw ocean trap and line fishery: spanner crab. (NSW Department of Primary Industries Fisheries) Available at https://www.dpi.nsw.gov.au/_data/assets/pdf_file/0004/1240735/NSW-Total-Allowable-Fishing-Committee-Spanner-Crab-Report-and-Determination-for-2020-21-fishing-period.pdf
- O'Connor MI, Bruno JF, Gaines SD, Halpern BS, Lester SE, Kinlan BP, Weiss JM (2007) Temperature control of larval dispersal and the implications for marine ecology, evolution, and conservation. *Proceedings of the National Academy of Sciences of the United States of America* **104**, 1266–1271. doi:10.1073/pnas.0603422104
- O'Neill MF, Campbell AB, Brown IW, Johnstone R (2010) Using catch rate data for simple cost-effective quota setting in the Australian spanner crab (*Ranina ranina*) fishery. *ICES Journal of Marine Science* **67**, 1538–1552. doi:10.1093/icesjms/fsq095
- Oke PR, Griffin DA, Schiller A, Matear RJ, Fiedler R, Mansbridge J, Lenton A, Cahill M, Chamberlain MA, Ridgway K (2013a) Evaluation of a near-global eddy-resolving ocean model. *Geoscientific Model Development* **6**, 591–615. doi:10.5194/gmd-6-591-2013
- Oke PR, Sakov P, Cahill ML, Dunn JR, Fiedler R, Griffin DA, Mansbridge J, Ridgway KR, Schiller A (2013b) Towards a dynamically balanced eddy-resolving ocean reanalysis: BRAN3. *Ocean Modelling* **67**, 52–70. doi:10.1016/j.ocemod.2013.03.008
- Oke PR, Roughan M, Cetina-Heredia P, Pilo GS, Ridgway KR, Rykova T, Archer MR, Coleman RC, Kerry CG, Rocha C, Schaeffer A, Vitarelli E (2019) Revisiting the circulation of the East Australian Current: its path, separation, and eddy field. *Progress in Oceanography* **176**, 102139. doi:10.1016/j.pocean.2019.102139
- Okubo A (1971) Oceanic diffusion diagrams. *Deep Sea Research and Oceanographic Abstracts* **18**, 789–802. doi:10.1016/0011-7471(71)90046-5
- Pecl GT, Araújo MB, Bell JD, Blanchard J, Bonebrake TC, Chen I-C, Clark TD, Colwell RK, Danielsen F, Evengård B, Falconi L, Ferrier S, Frusher S, Garcia RA, Griffis RB, Hobday AJ, Janion-Scheepers C, Jarzyna MA, Jennings S, Lenoir J, Linnertved HI, Martin VY, McCormack PC, McDonald J, Mitchell NJ, Mustonen T, Pandolfi JM, Pettorelli N, Popova E, Robinson SA, Scheffers BR, Shaw JD, Sorte CJB, Strugnell JM, Sunday JM, Tuanmu M-N, Vergés A, Villanueva C, Wernberg T, Wapstra E, Williams SE (2017) Biodiversity redistribution under climate change: impacts on ecosystems and human well-being. *Science* **355**, eaai9214. doi:10.1126/science.aai9214
- Poisot T (2011) The digitize package: extracting numerical data from scatterplots. *The R Journal* **3**, 25. doi:10.32614/RJ-2011-004
- Punt AE, Szuwalski C (2012) How well can FMSY and BMSY be estimated using empirical measures of surplus production? *Fisheries Research* **134–136**, 113–124. doi:10.1016/j.fishres.2012.08.014
- Putman NF, Mansfield KL (2015) Direct evidence of swimming demonstrates active dispersal in the sea turtle 'lost years'. *Current Biology* **25**, 1221–1227. doi:10.1016/j.cub.2015.03.014
- Queensland Government (2019) Queensland spanner crab fishery harvest strategy: 2020–2025, consultation draft. (Queensland Fisheries & Forestry) Available at https://www.daf.qld.gov.au/_data/assets/pdf_file/0008/1464209/qld-spanner-crab-fishery-harvest-strategy-consultation-draft-2020-2025.pdf [Verified 30 November 2021]
- Queensland Government (2021) Spanner crab management advice 2020. Available at <https://www.publications.qld.gov.au/dataset/b6593048-25b1-46c1-9feb-00a03d6501df/resource/e493b779-3c2b-4e2f-95b4-3a9ff778ad26/download/spanner-crab-management-advice.pdf>
- Roelofs A, Johnson D, McGilvray J (2021) Spanner crab (2020) *Ranina ranina* in status of australian fish stocks reports 2020. (Fisheries Research and Development Corporation) Available at <https://www.fish.gov.au/report/298-Spanner-Crab-2020>
- Roughan M, Macdonald HS, Baird ME, Glasby TM (2011) Modelling coastal connectivity in a Western Boundary Current: seasonal and inter-annual variability. *Deep-Sea Research – II. Topical Studies in Oceanography* **58**, 628–644. doi:10.1016/j.dsr2.2010.06.004
- Saenz-Agudelo P, Jones GP, Thorrold SR, Planes S (2011) Connectivity dominates larval replenishment in a coastal reef fish metapopulation. *Proceedings of the Royal Society of London – B. Biological Sciences* **278**, 2954–2961. doi:10.1098/rspb.2010.2780
- Scandol J, Kennelly S (2002) Using a fishery-independent survey to assess the status of a spanner crab *Ranina ranina* fishery: univariate analyses and biomass modelling. *Crustaceana* **75**, 13–39. doi:10.1163/156854002317373492
- Schaeffer A, Roughan M (2015) Influence of a western boundary current on shelf dynamics and upwelling from repeat glider deployments. *Geophysical Research Letters* **42**, 121–128. doi:10.1002/2014GL062260
- Schilling HT, Everett JD, Smith JA, Stewart J, Hughes JM, Roughan M, Kerry C, Suthers IM (2020) Multiple spawning events promote increased larval dispersal of a predatory fish in a western boundary current. *Fisheries Oceanography* **29**, 309–323. doi:10.1111/fog.12473
- Schilling HT, Hinchliffe C, Gillson JP, Miskiewicz AG, Suthers IM (2022) Coastal winds and larval fish abundance indicate a recruitment mechanism for southeast Australian estuarine fisheries. *Fisheries Oceanography* **31**, 40–55. doi:10.1111/fog.12561
- Skinner DG, Hill BJ (1987) Feeding and reproductive behaviour and their effect on catchability of the spanner crab *Ranina ranina*. *Marine Biology* **94**, 211–218. doi:10.1007/BF00392933
- Spencer DM, Brown IW, Lee SY, Lemckert CJ (2017) Physical oceanographic processes affecting catchability of spanner crab (*Ranina ranina*): a review. *Fisheries Research* **186**, 248–257. doi:10.1016/j.fishres.2016.09.005
- Spencer DM, Brown IW, Doubell MJ, Brown CJ, Rodriguez AR, Lee SY, Zhang H, Lemckert CJ (2019a) Bottom boundary layer cooling and wind-driven upwelling enhance the catchability of spanner crab (*Ranina ranina*) in South-East Queensland, Australia. *Fisheries Oceanography* **28**, 317–326. doi:10.1111/fog.12411
- Spencer DM, Brown IW, Doubell MJ, McGarvey R, Lee SY, Lemckert CJ (2019b) Bottom currents affect spanner crab catch rates in southern Queensland, Australia. *Marine and Coastal Fisheries* **11**, 248–257. doi:10.1002/mcf2.10073
- Spencer DM, Doubell MJ, Brown IW, Redondo Rodriguez A, Lee SY, Lemckert CJ (2019c) Environmental indices for spanner crab (*Ranina ranina*) catch rates depend on regional oceanographic features. *Estuarine, Coastal and Shelf Science* **228**, 106361. doi:10.1016/j.ecss.2019.106361

- Steven ADL, Baird ME, Brinkman R, Car NJ, Cox SJ, Herzfeld M, Hodge J, Jones E, King E, Margvelashvili N, Robillot C, Robson B, Schroeder T, Skerratt J, Tickell S, Tuteja N, Wild-Allen K, Yu J (2019) eReefs: an operational information system for managing the Great Barrier Reef. *Journal of Operational Oceanography* **12**, S12–S28. doi:10.1080/1755876X.2019.1650589
- Sun C, Feng M, Matear RJ, Chamberlain MA, Craig P, Ridgway KR, Schiller A (2012) Marine downscaling of a future climate scenario for Australian Boundary Currents. *Journal of Climate* **25**, 2947–2962. doi:10.1175/JCLI-D-11-00159.1
- Suthers IM, Young JW, Baird ME, Roughan M, Everett JD, Brassington GB, Byrne M, Condie SA, Hartog JR, Hassler CS, Hobday AJ, Holbrook NJ, Malcolm HA, Oke PR, Thompson PA, Ridgway K (2011) The strengthening East Australian Current, its eddies and biological effects: an introduction and overview. *Deep-Sea Research – II. Topical Studies in Oceanography* **58**, 538–546. doi:10.1016/j.dsr2.2010.09.029
- Szuwalski CS, Vert-Pre KA, Punt AE, Branch TA, Hilborn R (2015) Examining common assumptions about recruitment: a meta-analysis of recruitment dynamics for worldwide marine fisheries. *Fish and Fisheries* **16**, 633–648. doi:10.1111/faf.12083
- Taylor KE (2001) Summarizing multiple aspects of model performance in a single diagram. *Journal of Geophysical Research: Atmospheres* **106**, 7183–7192. doi:10.1029/2000JD900719
- Van Rossum G, Drake FL (2009) ‘Python 3 Reference Manual.’ (CreateSpace: Scotts Valley, CA, USA)
- van Sebille E, Griffies SM, Abernathy R, Adams TP, Berloff P, Biastoch A, Blanke B, Chassignet EP, Cheng Y, Cotter CJ, Deleersnijder E, Döös K, Drake HF, Drijfhout S, Gary SF, Heemink AW, Kjellsson J, Koszalka IM, Lange M, Lique C, MacGilchrist GA, Marsh R, Mayorga Adame CG, McAdam R, Nencioli F, Paris CB, Piggott MD, Polton JA, Rühls S, Shah SHAM, Thomas MD, Wang J, Wolfram PJ, Zanna L, Zika JD (2018) Lagrangian ocean analysis: fundamentals and practices. *Ocean Modelling* **121**, 49–75. doi:10.1016/j.ocemod.2017.11.008
- White JW, Carr MH, Caselle JE, Washburn L, Woodson CB, Palumbi SR, Carlson PM, Warner RR, Menge BA, Barth JA, Blanchette CA, Raimondi PT, Milligan K (2019) Connectivity, dispersal, and recruitment: connecting benthic communities and the coastal ocean. *Oceanography* **32**, 50–59. doi:10.5670/oceanog.2019.310
- Wu L, Cai W, Zhang L, Nakamura H, Timmermann A, Joyce T, McPhaden MJ, Alexander M, Qiu B, Visbeck M, Chang P, Giese B (2012) Enhanced warming over the global subtropical western boundary currents. *Nature Climate Change* **2**, 161–166. doi:10.1038/nclimate1353
- Yatsu A, Watanabe T, Ishida M, Sugisaki H, Jacobson LD (2005) Environmental effects on recruitment and productivity of Japanese sardine *Sardinops melanostictus* and chub mackerel *Scomber japonicus* with recommendations for management. *Fisheries Oceanography* **14**, 263–278. doi:10.1111/j.1365-2419.2005.00335.x

Data availability. All modelling code is available in the github repository (see https://github.com/DEHewitt/portunid_particle_tracking). The NSW commercial data are available only upon request from NSW Department of Primary Industries Fisheries.

Conflicts of interest. The authors declare that they have no conflicts of interest.

Declaration of funding. H. T. Schilling was supported by a NSW Government Research Attraction and Acceleration Program grant awarded to SIMS. D. E. Hewitt was supported by an Australian Government Research Training Program and a NSW DPIE top-up scholarship.

Acknowledgements. BRAN data are made freely available by CSIRO Bluelink and are supported by the Bluelink Partnership: a collaboration between the Australian Department of Defence, Bureau of Meteorology and CSIRO. This research includes computations using the computational cluster Katana supported by Research Technology Services at UNSW Sydney (see <https://doi.org/10.26190/669x-a286>). Mooring data were sourced from Australia’s Integrated Marine Observing System (IMOS); IMOS is enabled by the National Collaborative Research Infrastructure Strategy (NCRIS). We thank Professor Iain Suthers for support during this project. This is SIMS publication #296.

Author affiliations

^ASydney Institute of Marine Science, Chowder Bay Road, Mosman, NSW 2088, Australia.

^BCentre for Marine Science and Innovation, UNSW Sydney, Sydney, NSW 2052, Australia.

^CPort Stephens Fisheries Institute, New South Wales Department of Primary Industries, Locked Bag 1, Nelson Bay, NSW 2315, Australia.